

RESEARCH ARTICLE

Anthropogenic noise affects song structure in red-winged blackbirds (*Agelaius phoeniceus*)

Dalal Hanna¹, Gabriel Blouin-Demers¹, David R. Wilson^{2,*} and Daniel J. Mennill²

¹Department of Biology, University of Ottawa, Ottawa, ON, Canada, K1N 6N5 and ²Department of Biological Sciences, University of Windsor, 401 Sunset Avenue, Windsor, ON, Canada, N9B 3P4

*Author for correspondence (drwilson76@gmail.com)

Accepted 16 August 2011

SUMMARY

Anthropogenic noise can mask animal signals that are crucial for communicating information about food, predators and mating opportunities. In response to noise masking, signallers can potentially improve acoustic signal transmission by adjusting the timing, frequency or amplitude of their signals. These changes can be a short-term modification in response to transient noise or a long-term modification in response to chronic noise. An animal's ability to adapt to anthropogenic noise can be crucial to its success. In this study, we evaluated the effects of anthropogenic noise on the structure of red-winged blackbird song. First, we manipulated the presence of anthropogenic noise by experimentally broadcasting either silence or low-frequency white noise to subjects inhabiting quiet marshes located away from roadsides. Subjects exhibited increased signal tonality when temporarily exposed to low-frequency white noise, suggesting that red-winged blackbirds can alter their signals rapidly in response to sudden noise. Second, we compared songs produced in quiet marshes located away from roadsides with songs produced during quiet periods at roadside marshes that are normally noisy. This allowed us to test whether birds that are exposed to chronic anthropogenic noise exhibit altered song structure during temporarily quiet periods. Subjects residing in roadside marshes that are normally polluted with anthropogenic noise sang songs with increased tonality during quiet periods. Overall, our results show that anthropogenic noise influences the structure of birdsong. These effects should be considered in conservation and wildlife management.

Key words: anthropogenic noise, behavioural plasticity, birdsong, communication, conservation, red-winged blackbird, song structure.

INTRODUCTION

Acoustic communication is common among animals and is often essential for survival and reproduction (Bradbury and Vehrencamp, 1998; Catchpole and Slater, 1995). For example, acoustic signals can be used to defend territories (Sogge et al., 2007), attract mates (Catchpole and Slater, 1995; Saether, 2002), alert others to the presence of predators (Fallow and Magrath, 2010) and solicit food (Godfray, 1991; Ellis et al., 2009). For acoustic communication to occur, sounds must be detected and recognized by the receiver (Bradbury and Vehrencamp, 1998; Parks et al., 2011). The efficacy of communication can be influenced by many factors occurring at three distinct stages: signal production (e.g. signal structure, amplitude, directionality), signal transmission (e.g. humidity, ambient noise, topography) and signal perception (e.g. sensitivity, neuronal processing, appropriate behavioural response) (Bradbury and Vehrencamp, 1998; Wiley and Richards, 1982). Disruption of any one of these stages can impede communication.

Anthropogenic noise is a growing problem for many animals because it can mask their acoustic signals and lower their signal-to-noise ratio below a critical detection threshold (Halfwerk et al., 2010; Lohr et al., 2003; Pohl et al., 2009). Furthermore, as the human population grows and global urbanization continues, anthropogenic noise is likely to influence more animal populations. Although time has allowed animals to evolve several adaptive mechanisms to compensate for environmental noise, rapid anthropogenic changes could challenge the compensatory

mechanisms of many communication systems (Lengagne, 2008). Anthropogenic noise produced by roads, construction sites and airports, for instance, might affect breeding in species that are incapable of tolerating noise or unable to modify their signals to compensate for acoustic interference (Halfwerk et al., 2010; Slabbekoorn and Peet, 2003; Tyack, 2008). Therefore, evaluating responses to anthropogenic noise will become increasingly important for effective conservation and management of many animal species (Lengagne, 2008).

Several animals exhibit adaptations that increase their signal-to-noise ratio and thus diminish the effects of acoustic masking. There are at least five mechanisms by which animals can increase signal-to-noise ratio. First, animals can produce longer signals that are easier to detect (Brumm et al., 2004; Ey et al., 2009; Kirschel et al., 2009) or, as seen in certain anurans and katydid, they can adjust signal timing so that their signals do not overlap with conspecific, heterospecific or anthropogenic sounds (Fuller et al., 2007; Greenfield, 1994; Lengagne, 2008). Second, animals can increase signal amplitude during noisy periods (Pytte et al., 2003). This type of adjustment is known as the Lombard effect and has been observed in marine and terrestrial mammals, as well as birds (Brumm and Todt, 2002; Brumm et al., 2004; Brumm et al., 2009; Egnor and Hauser, 2006; Parks et al., 2011). Third, animals such as humpback whales (*Megaptera novaeangliae*) and house finches (*Carpodacus mexicanus*) can use a different type of signal that has better transmission properties during periods of high environmental noise

(Bermúdez-Cuamatzin et al., 2009; Dunlop et al., 2010). Fourth, animals can increase the tonality of their signals, as tonal signals transmit more efficiently than broadband signals through noisy environments (Lohr et al., 2003). For example, in habitats with high levels of natural noise (e.g. tropical forests), some species have been found to converge on songs that have primarily pure tones or narrow frequency bands (Dubois and Martens, 1984; Slabbekoorn and Smith, 2002). Fifth, animals can adjust the pitch of their signals to improve signal transmission (Laiolo, 2010). Green hylia (*Hylia prasina*), for example, avoid interference from insect sounds by decreasing the pitch of their songs (Kirschel et al., 2009), whereas certain anurans avoid interference from low-frequency stream noise by communicating with ultrasonic frequencies (Feng et al., 2006). In environments influenced by human activity, frequency shifts are usually towards higher frequencies, which minimize overlap with low-frequency anthropogenic noise (Slabbekoorn and Peet, 2003; Wood and Yezerinac, 2006).

Red-winged blackbirds, *Agelaius phoeniceus* (Linnaeus 1766), are an ideal species for studying the effects of anthropogenic noise on acoustic signalling because they commonly inhabit marshes located along noisy roads (Camp and Best, 1994). During the breeding season, red-winged blackbirds produce songs that vary in frequency from 1 to 5 kHz, and which typically consist of a series of introductory syllables followed by a broadband trill (Beletsky et al., 1980; Kroodsma and James, 1994). The trill is known to facilitate species recognition and is probably the part of the song that conveys the most meaningful information (Beletsky et al., 1980).

In this study, we examined the effects of anthropogenic noise on trill structure. In the first part of our study, we experimentally manipulated the presence of anthropogenic noise by broadcasting silence or low-frequency white noise to red-winged blackbirds in locations that normally do not experience traffic noise. This experimental approach allowed us to test for possible short-term plasticity in song structure. We predicted that trills would have a significantly longer duration, more tonal energy distribution, and a higher minimum frequency when they were produced during playback of low-frequency white noise. In the second part of our study, we compared songs that were produced during quiet periods by birds living near normally noisy roadsides with songs that were produced by birds living in locations that do not experience traffic noise. This allowed us to test for possible long-term changes in song structure. We predicted that the trill of red-winged blackbirds in marshes along noisy roadsides would have a significantly longer duration, more tonal energy distribution and a higher minimum frequency compared with the trill of individuals found in quiet marshes.

MATERIALS AND METHODS

Study area

We collected data on rainless mornings in May and June 2010 in the vicinity of the Queen's University Biological Station (44°34'N, 76°19'W), approximately 100 km southwest of Ottawa, ON, Canada.

Song structure during transient anthropogenic noise

We tested whether red-winged blackbirds from undisturbed marshes adjusted the structure of their trills in response to experimentally broadcasted low-frequency white noise. We presented a given subject with two sequential playback treatments over the course of approximately 6 min, while simultaneously recording its vocal response (recording details provided below). For each trial, we continued the first treatment until the subject sang between 3 and 7 songs. We commenced the second treatment immediately after

the first and continued it until the subject sang another 3–7 songs. Treatment order was randomized for each subject.

The two treatments were a silent control treatment in which we broadcast continuous silence, and an experimental noise treatment in which we broadcast continuous low-frequency white noise. To create the experimental noise treatment, we generated broad-spectrum white noise in Audition software (version 2.0; Adobe, San Jose, CA, USA) and then filtered the noise (bandpass filter, 0–1830 Hz) in Raven (version 1.4 Pro; Cornell Lab of Ornithology Bioacoustics Research Program, Ithaca, NY, USA). We chose an upper bandpass filter frequency of 1830 Hz because this frequency was lower than the minimum frequency of red-winged blackbird trills (minimum frequency of trills observed in our chronic noise experiment: mean \pm s.e.m., 2420 \pm 6 Hz; range, 2150–2790 Hz; see below). Thus, the experimental low-frequency white noise could be removed completely in the analysis without removing the lower frequencies of the trill, thereby enabling us to measure trill structure without those measurements being affected by the white noise. Furthermore, the average spectrum of our experimental noise approximated that of typical traffic noise, which has its dominant spectral energy below 1830 Hz (Cornillon and Keane, 1977; Halfwerk et al., 2010; Wood and Yezerinac, 2006).

Trials commenced when we identified a previously untested subject. To minimize the chances of testing the same bird on more than one occasion, we selected subjects by systematically circling a marsh. Individuals within a given marsh were distinguished by identifying distinct visual cues, such as plumage coloration, missing feathers and markings on the feet and bill (Bretagnolle et al., 1994; Lank and Dale, 2001). If subjects could not be distinguished reliably using visual cues, we separated trial locations by at least 200 m, which is further than the average territory diameter of red-winged blackbirds (Picman, 1987). Once a subject was identified, the observer approached the subject to within 5–30 m (23 \pm 1 m, mean \pm s.e.m.), placed a loudspeaker (Califone, model PA-285, frequency response 200–5000 Hz; San Fernando, CA, USA) 1–3 m from the observer, and connected the loudspeaker to a digital playback device (Apple iPod; Cupertino, CA, USA). The volume on the loudspeaker was set so that the amplitude of the experimental stimulus measured 89 dBC (C-weighted decibels) SPL (sound pressure level) at a distance of 1 m (measured with a RadioShack sound level meter, model 33-4050, slow response; Fort Worth, TX, USA). This amplitude is comparable to that used in another study [86 dBA SPL at a distance of 1 m (Verzijden et al., 2010)] and is typical of general highway traffic (mean \pm s.d., 72 \pm 18 dBA SPL at roadside, with 25% of all measurements >88 dBA SPL) (Lengagne, 2008) and heavy commercial truck traffic (10–90% range, 76–86 dBA SPL at a distance of 7.5 m) (Stephenson and Vulkan, 1968).

Subjects were audio-recorded throughout the trial using a digital recorder (Marantz PMD 660, WAVE format, 44.1 kHz sampling rate, 16 bits; Mahwah, NJ, USA) and a shotgun microphone (Audio-Technica, model 8015b; Stow, OH, USA) that was fitted with a foam windscreen.

Song structure following chronic anthropogenic noise

We compared red-winged blackbird songs produced in quiet marshes with those produced in marshes along a noisy roadside (Provincial Highway no. 15, ON, Canada) during temporarily quiet periods. We recorded 3–12 songs for each of 32 males distributed across four highway marshes (each separated by a minimum distance of 2 km), and an additional 3–12 songs for each of 32 males that were distributed across four non-highway marshes (each separated by a minimum of 2 km). To minimize the chances of recording the same

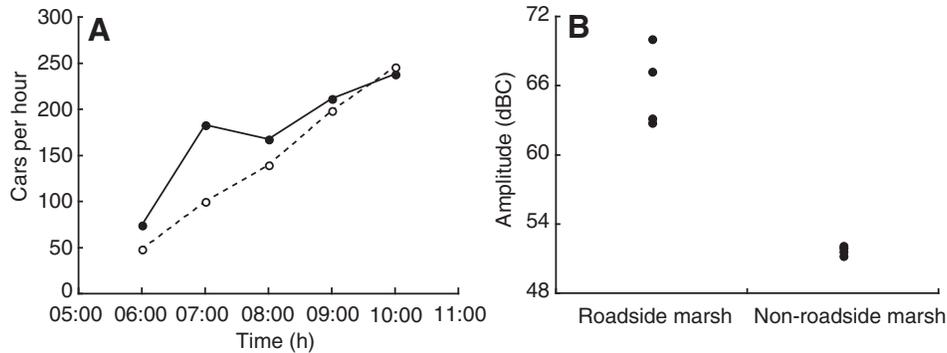


Fig. 1. Ambient noise at roadside and non-roadside marshes where red-winged blackbirds were recorded. (A) The mean daily incidence of vehicles on Highway 15 (ON, Canada) at the intersection of Road 8, between 06:00 h and 11:00 h, as measured by the Ministry of Transportation of Ontario during 1 week in April 2007. Filled circles with solid lines indicate weekday traffic ($N=5$ days) and open circles with dashed lines indicate weekend traffic ($N=2$ days). (B) Amplitude of ambient noise at four roadside recording locations is significantly higher than at four non-roadside recording locations on Highway 15. Each data point is a mean of sound pressure measurements that were obtained at 15 s intervals over a 5 min period.

bird on more than one occasion, recordings were made by systematically circling a marsh or by progressing north to south along the highway. Individuals within a given marsh were distinguished by identifying distinct visual cues. If individuals could not be distinguished reliably using visual cues, we separated recording locations by at least 200 m (Picman, 1987). We alternated our recordings daily between highway and non-highway marshes. Recordings were made using the recording apparatus described in the transient noise experiment above. The distance between the recorder and bird was estimated for each recording (mean \pm s.e.m., 20 ± 1 m; range, 4–50 m).

Data from the Ministry of Transportation of Ontario indicate high levels of traffic along Highway 15 (Fig. 1A): $0.3\text{--}4$ cars min^{-1} , with a speed limit of 80 km h^{-1} . To confirm that the level of ambient noise differed between highway and non-highway marshes, we measured the mean amplitude of ambient noise in each. Over the course of one morning, we went to eight typical recording locations distributed across four of our study sites (two highway and two non-highway). At each site, we measured the amplitude of ambient noise at 15 s intervals for a 5 min period (measured with the same sound level meter identified above). Roadside and non-roadside locations were alternated after every second trial to control for possible temporal variation in noise.

Measuring song structure

We analysed all recordings using Avisoft SASLab Pro (version 4.38; R. Specht, Berlin, Germany). First, we saved each song into a separate sound file that contained 1 s of silence both before and after the song (WAVE format, 16-bits, 44,100 Hz sampling rate). We then filtered all songs, regardless of the experiment or experimental treatment from which they were derived, with a 2 kHz high-pass filter, which removed natural and experimental background noise without removing components of the trill. Finally, we normalized songs to a peak amplitude of 0 dB and created a separate spectrogram for each song (1024 point fast Fourier transform, FFT; 87.5% overlap; Blackman window; 43 Hz frequency resolution; 2.9 ms temporal resolution).

We used Avisoft's 'automatic parameter measurement' feature (threshold setting, -15 dB; holdtime, 10 ms) to automatically select the trill of each song. Automatic detection always identified the end of the trill, but could not always distinguish between the introductory syllables and the onset of the trill. In these cases, we selected the onset of the trill manually by looking for the start of this syllable

on the spectrogram. With the trill selected, we collected multiple measurements: trill duration and the minimum frequency, maximum frequency and energy distribution (25%, 50% and 75% quartiles, as well as entropy) of the averaged frequency spectrum. Minimum frequency was considered the lowest frequency that exceeded -15 dB, whereas maximum frequency was considered the highest frequency that exceeded -15 dB. Quartiles were defined as the frequencies that divided the spectrum into four parts that were equal in terms of spectral energy. Entropy was defined as the randomness of sound: a pure tone has an entropy value near 0, whereas white noise has an entropy value near 1. For roadside recordings, we did not analyse any songs for which we could hear traffic noise in the background or see traffic noise in the sound spectrogram.

A concern in the transient noise experiment was that the broadcasted low-frequency white noise could have directly affected the measurements of song structure (Verzijden et al., 2010). This seems unlikely, however, as our 2 kHz high-pass filter should have completely removed the low-frequency white noise without affecting trill structure. Nevertheless, we conducted a validation procedure similar to that used by Verzijden and colleagues (Verzijden et al., 2010) to ensure that our measurements were not affected directly by the experimental treatment. To achieve this, we replicated our transient anthropogenic noise experiment but, instead of recording live subjects, we recorded simulated subjects that could not have changed the structure of their songs between the two experimental treatments. Thus, any treatment effects detected by our measurements in this validation procedure would be artifacts caused by the experimental broadcast of low-frequency white noise. In our validation procedure, we simulated 20 subjects by broadcasting pre-recorded songs from 20 different red-winged blackbirds at 20 different locations in a marsh habitat. We broadcast songs at natural amplitude (85 dBC SPL at a distance of 1 m) (Patricelli et al., 2007) through a loudspeaker (Califone, model PA-285) placed 17–30 m from the observer. Simultaneously, we broadcast silence (i.e. the control treatment described above) or noise (i.e. the experimental treatment described above) through a second loudspeaker (Califone, model PA-285) placed 1–3 m from the observer, and then repeated the procedure using the opposite treatment. Consequently, the only difference between treatments was the presence or absence of experimental low-frequency white noise (i.e. song structure was held constant between the treatments). This allowed us to determine whether the broadcast of noise influenced our measurements of birdsong.

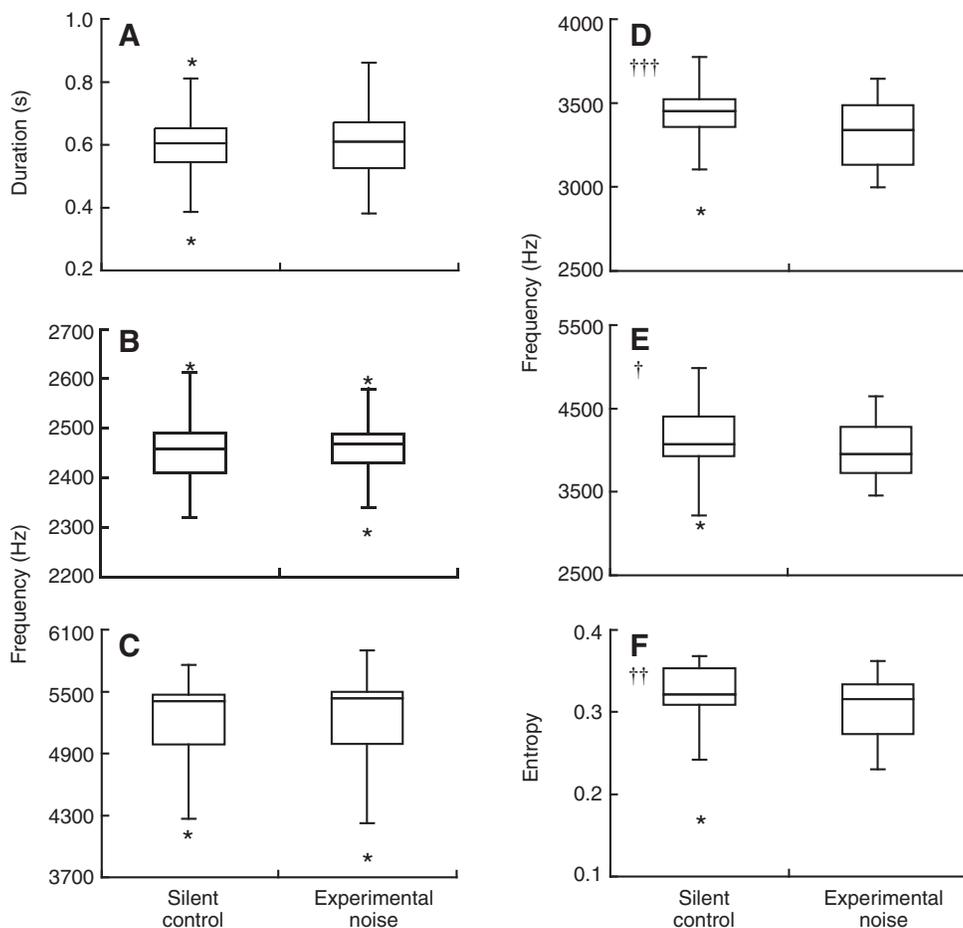


Fig. 2. Variation in the structure of red-winged blackbird songs during the transient anthropogenic noise experiment. Each subject ($N=20$) was presented with a silent control treatment, in which we broadcast silence, and an experimental noise treatment, in which we broadcast low-frequency white noise. Box plots show the median (black line), inter-quartile range (box), minimum value within 1.5 inter-quartile ranges of the lower quartile (lower whisker), maximum value within 1.5 inter-quartile ranges of the upper quartile (upper whisker), and minimum and maximum outliers (asterisks) for six song parameters measured from the trill component of the birds' songs: (A) duration, (B) minimum frequency, (C) maximum frequency, (D) 50% quartile, (E) 75% quartile and (F) entropy. For each parameter, each data point is the mean of all songs produced by a given individual. Statistically significant differences are indicated by daggers in the top left of each panel ($\dagger P \leq 0.05$, $\dagger\dagger P \leq 0.01$ and $\dagger\dagger\dagger P \leq 0.005$).

Statistical analyses

The 25% quartile was highly correlated with the 50% and 75% quartiles, so we removed it from the analysis to avoid conducting independent analyses on non-independent response variables. We constructed separate linear mixed models for each of the measured song characteristics by entering them as response variables. In each model, we included treatment (silence *versus* low-frequency white noise in the transient noise experiment) or habitat type (roadside *versus* non-roadside in the chronic noise comparison) as a factor with fixed effects and the distance between the subject and recorder as a covariate with fixed effects. We accounted for repeated measurements of the same individual by including 'individual' as a subject factor with random effects.

Assumptions of normality and homoscedasticity were verified by inspecting histograms. The entropy variable in the chronic noise experiment was not normally distributed but was corrected with a cubic transformation prior to analysis. Results were considered statistically significant when $P \leq 0.05$, and non-significant trends were considered when $0.05 < P \leq 0.10$. Means are reported ± 1 standard error and distance effects are described using unstandardized regression coefficients ($B \pm \text{s.e.}$). All statistical analyses were conducted in PASW for Mac (version 19.0; Armonk, NY, USA).

RESULTS

Song structure during transient anthropogenic noise

We analysed 140 songs from 20 male red-winged blackbirds (69 songs during silent control treatment; 71 songs during experimental noise treatment). Experimental noise treatment did not affect trill

duration (linear mixed model: treatment effect, $F_{1,118}=0.009$, $P=0.927$; distance effect, $F_{1,100}=0.791$, $P=0.376$; $B \pm \text{s.e.}=0.002 \pm 0.002$), minimum frequency (treatment effect, $F_{1,116}=0.633$, $P=0.428$; distance effect, $F_{1,47}=0.189$, $P=0.666$; $B \pm \text{s.e.}=-0.700 \pm 1.610$) or maximum frequency (treatment effect, $F_{1,119}=0.231$, $P=0.632$; distance effect, $F_{1,52}=2.413$, $P=0.126$; $B \pm \text{s.e.}=-15.308 \pm 9.854$). Experimental noise treatment did affect entropy, however, with trills produced during playback of low-frequency white noise being significantly more tonal than trills produced during the silent control (linear mixed model: treatment effect, $F_{1,118}=7.293$, $P=0.008$; distance effect, $F_{1,73}=4.746$, $P=0.033$; $B \pm \text{s.e.}=-0.001 \pm 0.001$). The increased tonality during the experimental noise treatment was due to the spectral energy of the trill being concentrated in lower frequencies. Specifically, the 50% and 75% quartiles were significantly lower during the experimental noise treatment than during the silent control treatment (linear mixed models: treatment effect for 50% quartile, $F_{1,118}=9.789$, $P=0.002$; distance effect for 50% quartile, $F_{1,67}=0.785$, $P=0.379$; $B \pm \text{s.e.}=-3.259 \pm 3.677$; treatment effect for 75% quartile: $F_{1,117}=3.949$, $P=0.049$; distance effect for 75% quartile, $F_{1,60}=5.201$, $P=0.026$; $B \pm \text{s.e.}=-14.494 \pm 6.355$; Fig. 2).

We conducted a validation procedure to ensure that our measurements of song structure were not influenced by the presence of low-frequency white noise. We analysed 138 songs from 20 simulated male red-winged blackbirds (69 songs during playback of the silent control stimulus and the same 69 songs during playback of experimental noise). In marked contrast to our analysis of live subjects, experimental noise did not affect any of the temporal or

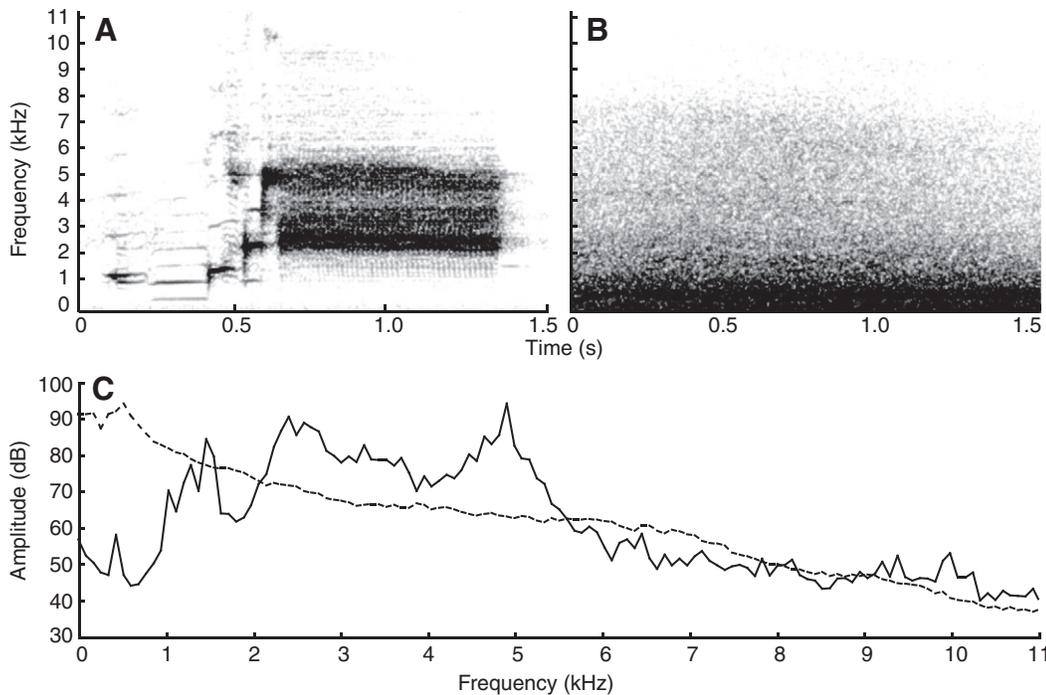


Fig. 3. The masking effect of traffic noise on a typical red-winged blackbird song. Spectrograms of the song (A) and the traffic noise (B) were produced using a 1024 point fast Fourier transform (FFT), 87.5% overlap, and a Blackman window, which resulted in a frequency resolution of 43 Hz and a temporal resolution of 2.9 ms. (C) Averaged power spectrum of the red-winged blackbird song (solid line) and the traffic noise (dashed line).

spectral characteristics of the simulated songs (linear mixed models: treatment effect for duration, $F_{1,116}=0.004$, $P=0.951$; distance effect for duration, $F_{1,18}=6.639$, $P=0.019$; $B \pm \text{s.e.} = 0.009 \pm 0.004$; treatment effect for minimum frequency, $F_{1,117}=0.059$, $P=0.808$; distance effect for minimum frequency, $F_{1,18}=2.128$, $P=0.162$; $B \pm \text{s.e.} = 3.755 \pm 2.574$; treatment effect for maximum frequency, $F_{1,117}=0.061$, $P=0.805$; distance effect for maximum frequency, $F_{1,18}=0.013$, $P=0.910$; $B \pm \text{s.e.} = -4.178 \pm 36.347$; treatment effect for entropy, $F_{1,121}=0.591$, $P=0.444$; distance effect for entropy, $F_{1,17}=0.611$, $P=0.445$; $B \pm \text{s.e.} = -0.001 \pm 0.002$; treatment effect for 50% quartile, $F_{1,117}=0.718$, $P=0.399$; distance effect for 50% quartile, $F_{1,18}=0.026$, $P=0.874$; $B \pm \text{s.e.} = -1.667 \pm 10.396$; treatment effect for 75% quartile, $F_{1,117}=0.678$, $P=0.412$; distance effect for 75% quartile, $F_{1,18}=0.600$, $P=0.449$; $B \pm \text{s.e.} = -10.084 \pm 13.023$). This demonstrates that the changes observed in the songs of the subjects of our playback experiment were not an artifact of the experimental noise treatment.

Song structure following chronic anthropogenic noise

The mean amplitude of ambient noise at roadside locations (65.8 ± 1.8 dB SPL) was higher than the ambient noise observed at non-roadside locations (51.7 ± 0.1 dB SPL), owing to a minimum of one car passing every 2 min on the highway (Fig. 1B). In addition, the power spectrum of a typical red-winged blackbird trill overlaps that of typical traffic noise, thereby revealing the potential for acoustic masking (Fig. 3).

We recorded 436 songs from 64 male red-winged blackbirds (32 in roadside marshes, 32 in non-roadside marshes). We excluded 92 songs because they contained obvious traffic noise, resulting in 344 songs from 63 individuals for our comparison of roadside *versus* non-roadside songs (32 individuals from non-roadside marshes, 31 individuals from roadside marshes). In general, the results were similar to those from the transient anthropogenic noise experiment. Habitat type (i.e. roadside *versus* non-roadside) did not affect trill duration (linear mixed model: treatment effect, $F_{1,60}=0.337$, $P=0.564$; distance effect, $F_{1,318}=2.058$, $P=0.152$; $B \pm$

$\text{s.e.} = 0.001 \pm 0.001$) or minimum frequency (treatment effect, $F_{1,59}=2.751$, $P=0.103$; distance effect, $F_{1,148}=1.174$, $P=0.280$; $B \pm \text{s.e.} = 0.938 \pm 0.866$). Habitat type did affect trill entropy, however, with trills recorded in roadside marshes being significantly more tonal than trills recorded in non-roadside marshes (linear mixed model: treatment effect, $F_{1,61}=6.027$, $P=0.017$; distance effect, $F_{1,324}=12.470$, $P<0.001$; $B \pm \text{s.e.} = -0.0003 \pm 0.0001$; Fig. 4). The increased tonality of trills recorded along noisy roadsides was again due to the spectral energy of those trills being concentrated at lower frequencies. The maximum frequency and the 75% quartile were both significantly lower for trills recorded along noisy roadsides than for trills recorded in non-roadside marshes (linear mixed models: treatment effect for maximum frequency, $F_{1,52}=10.273$, $P=0.002$; distance effect for maximum frequency, $F_{1,117}=0.354$, $P=0.553$; $B \pm \text{s.e.} = -2.235 \pm 3.757$; treatment effect for 75% quartile, $F_{1,60}=4.712$, $P=0.034$; distance effect for 75% quartile, $F_{1,148}=0.941$, $P=0.334$; $B \pm \text{s.e.} = -3.057 \pm 3.152$; Fig. 4). The 50% quartile also tended to be lower for trills recorded in noisy roadside marshes than for trills recorded in quiet non-roadside marshes, though the effect was not statistically significant (linear mixed model: treatment effect, $F_{1,61}=3.477$, $P=0.067$; distance effect, $F_{1,211}=1.785$, $P=0.183$; $B \pm \text{s.e.} = -2.158 \pm 1.615$; Fig. 4).

DISCUSSION

We tested whether anthropogenic noise affects the structure of red-winged blackbird song using two complementary approaches. First, we used playback to manipulate the presence of anthropogenic noise in pristine marshes, far from busy roadsides. Subjects exposed to transient experimental low-frequency white noise immediately adjusted the structure of their songs by increasing signal tonality, which demonstrates a high degree of behavioural plasticity in the singing behaviour of this species. The change in tonality was not a measurement artifact created by the presence of experimental noise, as we did not observe the same result when we replaced live subjects with simulated subjects that could not alter the structure of their songs. In the second part of our study, we compared song structure

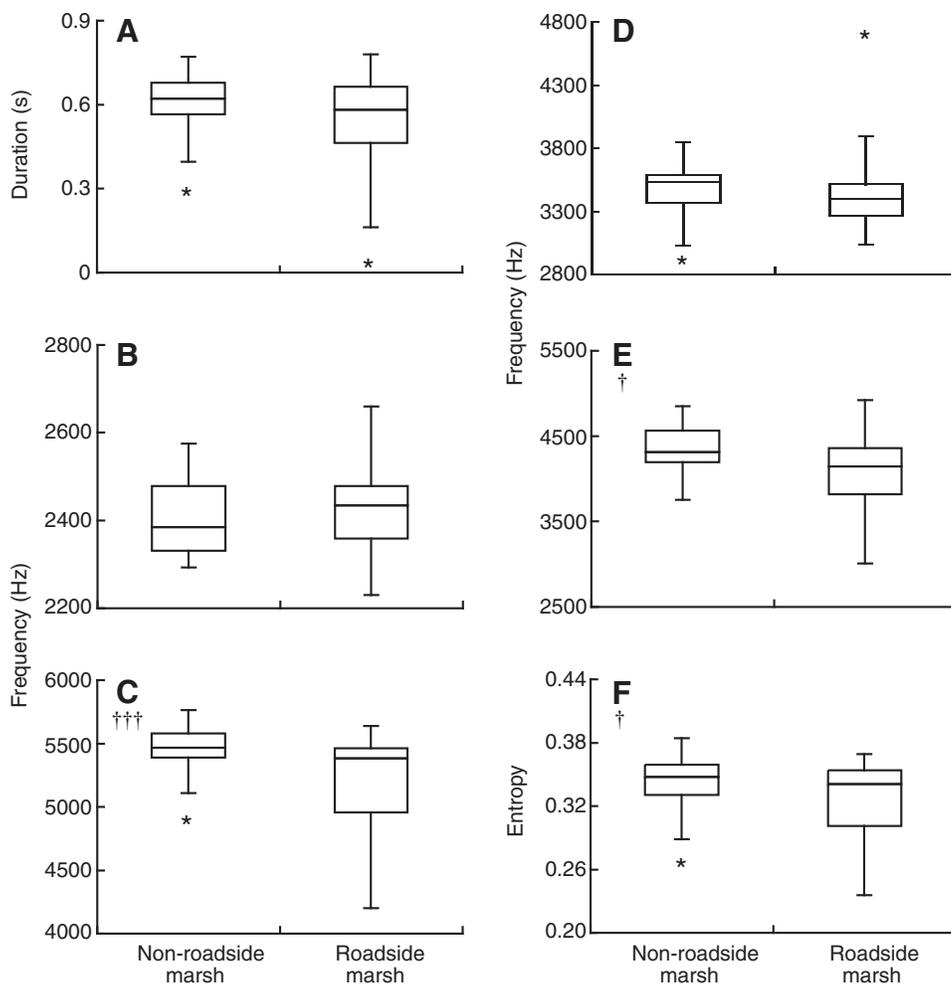


Fig. 4. Variation in the structure of red-winged blackbird songs for subjects that were recorded in quiet non-roadside marshes ($N=32$) and in noisy roadside marshes when traffic was temporarily absent ($N=31$). Box plots show the median (black line), inter-quartile range (box), minimum value within 1.5 inter-quartile ranges of the lower quartile (lower whisker), maximum value within 1.5 inter-quartile ranges of the upper quartile (upper whisker), and minimum and maximum outliers (asterisks) for six parameters measured from the trill component of birds' songs: (A) duration (B) minimum frequency, (C) maximum frequency, (D) 50% quartile, (E) 75% quartile and (F) entropy. For each parameter, each data point is the mean of all songs produced by a given individual. Statistically significant differences are indicated by daggers in the top left of each panel († $P \leq 0.05$, †† $P \leq 0.01$ and ††† $P \leq 0.005$).

between populations that differ naturally in terms of chronic anthropogenic noise. Subjects that live near a busy highway, where they encounter chronic anthropogenic noise, produced songs with increased tonality during temporarily quiet periods, demonstrating that long-term exposure to chronic anthropogenic noise has lasting effects on red-winged blackbird song.

Our results demonstrate an overall increase in signal tonality caused by an emphasis of lower frequencies when red-winged blackbirds are exposed to anthropogenic noise. This effect has not been observed previously in response to anthropogenic noise, though a similar effect has been observed in response to natural noise. Large-billed leaf-warblers (*Phylloscopus magnirostris*) and several species of frog (genus *Rana*, subgenus *Paa*) live near the noisy torrents of the Nepal Himalayas. Compared with congeners living in habitats devoid of torrents, these species produce signals with increased tonality (Dubois and Martens, 1984). The reasons for increasing signal tonality in noisy environments, however, remain unclear. One possible reason is that higher frequencies attenuate more rapidly than lower frequencies (Bradbury and Vehrencamp, 1998; Forest, 1994). Re-distributing spectral energy from higher frequencies to lower frequencies could therefore enhance signal transmission in noisy environments. This strategy could be particularly effective for red-winged blackbirds living along roadsides because the lower frequencies contained in their trills (2.75–5.0 kHz) generally exceed the higher frequencies produced by traffic (traffic noise is concentrated below 3 kHz) (Cornillon and

Keane, 1977; Halfwerk et al., 2010; Wood and Yezerinac, 2006). Another possible reason for increasing tonality in noisy environments is that narrow-band signals are easier than broadband signals for receivers to discriminate from background noise (Lohr et al., 2003). This is because the energy in a tonal signal is concentrated in the frequency domain, which increases its signal-to-noise ratio in the part of the spectrum containing the signal (Lohr et al., 2003). It is unclear whether tonal signals are also easier to detect when the signal and background noise occupy completely different parts of the frequency spectrum, as was the case in our transient noise experiment. Research on humans, however, suggests that the noise stimulus could still have disrupted the detection of trills and given subjects incentive to modify trill structure, even though the frequencies of the noise stimulus did not overlap the frequencies of the trill (Martin and Pickett, 1970; Berglund et al., 1996; Vinnik et al., 2011). Finally, it is also possible that increased tonality is not an adaptation to noisy environments but a non-adaptive side-effect of another mechanism such as the Lombard effect (Lu and Cooke, 2009).

There are several strategies that animals can use to improve signal detectability in noisy environments, other than increasing signal tonality, yet we did not detect any of these alternative strategies in our study. For example, prolonging signals is a common strategy for increasing signal detectability (Brumm et al., 2004; Nemeth et al., 2006; Warren et al., 2006), yet the trill duration in our study did not differ between all treatments. It is possible that red-winged

blackbirds are unable to prolong their trills or that there are energetic costs associated with prolonging trills that outweigh the potential benefits (Fernández-Juricic et al., 2005). Another strategy for increasing signal detectability in noisy environments is to increase signal amplitude. Future studies using an amplitude-calibrated recording system would facilitate quantitative measurement of whether red-winged blackbirds modify the amplitude of their songs. Many species improve signal detectability by increasing the minimum frequency of their songs, which, because of the low-frequency nature of anthropogenic noise, can reduce masking effects (e.g. Bermúdez-Cuamatzin et al., 2009; Kirschel et al., 2009; Ripmeester et al., 2010; Verzijden et al., 2010). In our study, the minimum frequency of trills did not differ between treatments, perhaps because the frequency range of red-winged blackbird trills (2.75–5.0 kHz) exceeded the frequencies contained in our experimental low-frequency white noise (0–1.83 kHz) and exceeded the majority of frequencies produced by traffic (again, traffic noise is concentrated below 3 kHz) (Cornillon and Keane, 1977; Halfwerk et al., 2010; Wood and Yezerinac, 2006). Finally, some species increase signal detectability in noisy environments by producing an alternative type of signal with better transmission properties (e.g. Bermúdez-Cuamatzin et al., 2009; Dunlop et al., 2010). Red-winged blackbirds have only a single song type, however, so this strategy is not an option for this species.

Our results demonstrate that red-winged blackbirds modify the structure of their songs in the presence of anthropogenic noise. Although such modifications presumably facilitate signal transmission and detection, the modifications can also impose significant costs on the signaller (Parks et al., 2011; Tyack, 2008). For example, females could have difficulty recognizing modified songs, or could be less attracted to males that produce modified songs (Patricelli and Blickley, 2006). Structural modifications to song could also have consequences that are manifested during intrasexual interactions. For example, males that have adjusted their signals could be perceived by other males as being less threatening, thereby compromising their ability to defend a territory (Patricelli and Blickley, 2006). If males with altered songs are unable to attract a mate or defend a territory, then populations affected by anthropogenic noise may gradually disappear. Alternatively, if females prefer the songs of their local males, then roadside and non-roadside populations may become reproductively isolated, which would constitute the first steps of speciation (Slabbekoorn and Peet, 2003; Warren et al., 2006; Wood and Yezerinac, 2006). Conservation authorities should therefore consider a number of strategies for reducing anthropogenic noise in protected areas and areas with species at risk (Kam-Wah, 2005; Barber et al., 2010).

Future studies should focus on how the structural changes observed in the current study affect signal transmission and signal perception. For example, does increased signal tonality improve signal transmission in a noisy environment? Do receivers discriminate between more tonal and less tonal songs, and, if so, do these differences in song structure affect the signaller's survival and reproductive success (Nemeth and Brumm, 2010; Verzijden et al., 2010)? The results we present here show that red-winged blackbirds modify their acoustic mating signals in response to anthropogenic noise, and that animals living in areas with chronic exposure to high noise levels show long-term changes to their song structure.

ACKNOWLEDGEMENTS

We thank M. Jackson, N. Asi and the staff at Queen's University Biological Station for technical and logistical assistance, M. Conboy for help with mapping and

weather data, S. Sadaka, A. Hanna and L.-P. Robillard for field assistance, and two anonymous reviewers for insightful comments on an earlier version of our manuscript.

FUNDING

Funding was provided by the American Ornithologists' Union, the Natural Sciences and Engineering Research Council of Canada, the Ontario Ministry of Research and Innovation, and the Canada Foundation for Innovation.

REFERENCES

- Barber, J. R., Crooks, K. R. and Fristrup, K. M. (2010). The costs of chronic noise exposure for terrestrial organisms. *Trends Ecol. Evol.* **25**, 180–189.
- Beletsky, L. D., Chao, S. and Smith, D. G. (1980). An investigation of song-based species recognition in the red-winged blackbird (*Abelauis pheoniceus*). *Behaviour* **73**, 189–203.
- Berglund, B., Hassmen, P. and Job, R. F. S. (1996). Sources and effects of low-frequency noise. *J. Acoust. Soc. Am.* **99**, 2985–3002.
- Bermúdez-Cuamatzin, E., Rios-Chelén, A. A., Gil, D. and García, C. M. (2009). Strategies of song adaptation to urban noise in the house finch: syllable pitch plasticity or differential syllable use? *Behaviour* **146**, 1269–1286.
- Bradbury, J. W. and Vehrencamp, S. L. (1998). *Principles of Animal Communication*. Sunderland: Sinauer Associates.
- Bretagnolle, V., Thibault, J. C. and Dominic, J. M. (1994). Field identification of individual ospreys using head marking pattern. *J. Wildl. Manage.* **58**, 175–178.
- Brumm, H. and Todt, D. (2002). Noise-dependent song amplitude regulation in a territorial songbird. *Anim. Behav.* **63**, 891–897.
- Brumm, H., Voss, K., Kollmer, I. and Todt, D. (2004). Acoustic communication in noise: regulation of call characteristics in a New World monkey. *J. Exp. Biol.* **207**, 443–448.
- Brumm, H., Schmidt, R. and Schrader, L. (2009). Noise-dependent vocal plasticity in domestic fowl. *Anim. Behav.* **78**, 741–746.
- Camp, M. and Best, L. B. (1994). Nest density and nesting success of birds in roadsides adjacent to rowcrop fields. *Am. Midl. Nat.* **131**, 347–358.
- Catchpole, C. K. and Slater, P. J. B. (1995). *Bird Song: Biological Themes and Variations*. New York: Cambridge University Press.
- Cornillon, P. C. and Keane, M. A. (1977). Simple model for simulating traffic noise spectra. *J. Acoust. Soc. Am.* **61**, 739–743.
- Dubois, A. and Martens, J. (1984). A case of possible vocal convergence between frogs and a bird in Himalayan torrents. *J. Ornithol.* **125**, 455–463.
- Dunlop, R. A., Cato, D. H. and Noad, M. J. (2010). Your attention please: increasing ambient noise levels elicits a change in communication behaviour in humpback whales (*Megaptera novaeangliae*). *Proc. R. Soc. Lond. B* **277**, 2521–2529.
- Egnor, S. E. R. and Hauser, M. D. (2006). Noise-induced vocal modulation in cotton-top tamarins (*Saguinus oedipus*). *Am. J. Primatol.* **68**, 1183–1190.
- Ellis, J. M. S., Langen, T. A. and Berg, E. C. (2009). Signalling for food and sex? Begging by reproductive female white-throated magpie-jays. *Anim. Behav.* **78**, 615–623.
- Ey, E., Rahn, C., Hammerschmidt, K. and Fischer, J. (2009). Wild female olive baboons adapt their grunt vocalizations to environmental conditions. *Ethology* **115**, 493–503.
- Fallow, P. M. and Magrath, R. D. (2010). Eavesdropping on other species: mutual interspecific understanding of urgency information in avian alarm calls. *Anim. Behav.* **79**, 411–417.
- Feng, A. S., Narins, P. M., Xu, C. H., Lin, W. Y., Yu, Z. L., Qiu, Q., Xu, Z. M. and Shen, J. X. (2006). Ultrasonic communication in frogs. *Nature* **440**, 333–336.
- Fernández-Juricic, E., Poston, R., De Collibus, K., Morgan, T., Bastain, B., Martin, C., Jones, K. and Termino, R. (2005). Microhabitat selection and singing behavior patterns of male house finches (*Carpodacus mexicanus*) in urban parks in a heavily urbanized landscape in the western U.S. *Urban Habitats* **3**, 49–69.
- Forrest, T. G. (1994). From sender to receiver – propagation and environmental effects on acoustic signals. *Am. Zool.* **34**, 644–654.
- Fuller, R. A., Warren, P. H. and Gaston, K. J. (2007). Daytime noise predicts nocturnal singing in urban robins. *Biol. Lett.* **3**, 368–370.
- Godfray, H. C. J. (1991). Signaling of need by offspring to their parents. *Nature* **352**, 328–330.
- Greenfield, M. D. (1994). Cooperation and conflict in the evolution of signal interactions. *Annu. Rev. Ecol. Syst.* **25**, 97–126.
- Halfwerk, W., Holleman, L. J. M., Lessells, C. K. M. and Slabbekoorn, H. (2010). Negative impact of traffic noise on avian reproductive success. *J. Appl. Ecol.* **48**, 210–219.
- Kam-Wah, L. (2005). Traffic noise mitigation in Hong Kong. MSc thesis, The University of Hong Kong, Hong Kong, China.
- Kirschel, A. N. G., Blumstein, D. T., Cohen, R. E., Buermann, W., Smith, T. B. and Slabbekoorn, H. (2009). Birdsong tuned to the environment: green hylia song varies with elevation, tree cover, and noise. *Behav. Ecol.* **20**, 1089–1095.
- Kroodsma, D. E. and James, F. C. (1994). Song variation within and among populations of red-winged blackbirds. *Wilson Bull.* **106**, 156–162.
- Laio, P. (2010). The emerging significance of bioacoustics in animal species conservation. *Biol. Conserv.* **143**, 1635–1645.
- Lank, D. B. and Dale, J. (2001). Visual signals for individual identification: the silent 'song' of Ruffs. *Auk* **118**, 759–765.
- Lengagne, T. (2008). Traffic noise affects communication behaviour in a breeding anuran, *Hyla arborea*. *Biol. Conserv.* **141**, 2023–2031.
- Lohr, B., Wright, T. F. and Dooling, R. J. (2003). Detection and discrimination of natural calls in masking noise by birds: estimating the active space of a signal. *Anim. Behav.* **65**, 763–777.

- Lu, Y. and Cooke, M.** (2009). Speech production modifications produced in the presence of low-pass and high-pass filtered noise. *J. Acoust. Soc. Am.* **126**, 1495-1499.
- Martin, E. S. and Pickett, J. M.** (1970). Sensorineural hearing loss and upward spread of masking. *J. Speech Hear. Res.* **13**, 426-437.
- Nemeth, E. and Brumm, H.** (2010). Birds and anthropogenic noise: are urban songs adaptive? *Am. Nat.* **176**, 465-475.
- Nemeth, E., Dabelsteen, T., Pedersen, S. B. and Winkler, H.** (2006). Rainforests as concert halls for birds: are reverberations improving sound transmission of long song elements? *J. Acoust. Soc. Am.* **119**, 620-626.
- Parks, S. E., Johnson, M., Nowacek, D. and Tyack, P. L.** (2011). Individual right whales call louder in increased environmental noise. *Biol. Lett.* **7**, 33-35.
- Patricelli, G. L. and Blickley, J. L.** (2006). Avian communication in urban noise: causes and consequences of vocal adjustment. *Auk* **123**, 639-649.
- Patricelli, G. L., Dantzker, M. S. and Bradbury, J. W.** (2007). Differences in acoustic directionality among vocalizations of the male red-winged blackbird (*Agelaius phoeniceus*) are related to function in communication. *Behav. Ecol. Sociobiol.* **61**, 1099-1110.
- Picman, J.** (1987). Territory establishment, size, and tenacity by male red-winged blackbirds. *Auk* **104**, 405-412.
- Pohl, N. U., Slabbekoorn, H., Klump, G. M. and Langemann, U.** (2009). Effects of signal features and environmental noise on signal detection in the great tit, *Parus major*. *Anim. Behav.* **78**, 1293-1300.
- Pytte, C. L., Rusch, K. M. and Ficken, M. S.** (2003). Regulation of vocal amplitude by the blue-throated hummingbird, *Lampornis clemenciae*. *Anim. Behav.* **66**, 703-710.
- Ripmeester, E. A. P., Mulder, M. and Slabbekoorn, H.** (2010). Habitat-dependent acoustic divergence affects playback response in urban and forest populations of the European blackbird. *Behav. Ecol.* **21**, 876-883.
- Saether, S. A.** (2002). Female calls in lek-mating birds: indirect mate choice, female competition for mates, or direct mate choice? *Behav. Ecol.* **13**, 344-352.
- Slabbekoorn, H. and Peet, M.** (2003). Birds sing at a higher pitch in urban noise. *Nature* **424**, 267.
- Slabbekoorn, H. and Smith, T. B.** (2002). Habitat-dependent song divergence in the little greenbul: an analysis of environmental selection pressures on acoustic signals. *Evolution* **56**, 1849-1858.
- Sogge, M. K., Koronkiewicz, T. J., van Riper, C. and Durst, S. L.** (2007). Willow flycatcher nonbreeding territory defense behavior in Costa Rica. *Condor* **109**, 475-480.
- Stephenson, R. J. and Vulkan, G. H.** (1968). Traffic noise. *J. Sound Vib.* **7**, 247-262.
- Tyack, P. L.** (2008). Implications for marine mammals of large-scale changes in the marine acoustic environment. *J. Mammal.* **89**, 549-558.
- Verzijden, M. N., Ripmeester, E. A. P., Ohms, V. R., Snelderwaard, P. and Slabbekoorn, H.** (2010). Immediate spectral flexibility in singing chiffchaffs during experimental exposure to highway noise. *J. Exp. Biol.* **213**, 2575-2581.
- Vinnik, E., Itskov, P. M. and Balaban, E.** (2011). Individual differences in sound-in-noise perception are related to the strength of short-latency neural responses to noise. *PLoS ONE* **6**, e17266.
- Warren, P. S., Katti, M., Ermann, M. and Brazel, A.** (2006). Urban bioacoustics: it's not just noise. *Anim. Behav.* **71**, 491-502.
- Wiley, R. H. and Richards, D. G.** (1982). Adaptations for acoustic communication in birds: sound transmission and signal detection. In *Acoustic Communication in Birds*, Vol. 1 (ed. D. E. Kroodsma and E. H. Miller), pp. 131-181. New York: Academic Press.
- Wood, W. E. and Yezerinac, S. M.** (2006). Song sparrow (*Melospiza melodia*) song varies with urban noise. *Auk* **123**, 650-659.